

IMPACT OF GENETICALLY MODIFIED PLANTS ON THE ENVIRONMENT

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ABSTRACT: Application technologies are considered genetic engineering in agriculture with the introduction of genetically modified (GM) plants are constantly creating new opportunities to increase crop production and to solve problems in agriculture, such as diseases, pests, weeds, abiotic stresses, and nutritional limitations. These are formed, and the plants having new properties that enable their use in pharmaceutical manufacturing. As the GM plant introductions to various locations characterized by a variety of ecosystems, types of agriculture, biodiversity and agriculturally practice, it is necessary to the scientific understanding of the effects of growing GM plants on the environment, which will provide security and environmental sustainability. The most important research she focused on the invasiveness of the GM plant, vertical and horizontal gene transfer, the impact on biological diversity and on other products. These tests are very complex multi, inter and transdisciplinary. This article discusses some of the most important problems related to the application of the technology of genetic engineering in agriculture and the introduction of GM plants into the environment, such as plant protection, ecological effects of horizontal gene transfer, biodiversity, stress, the effects of land, etc. There is a clear need to further investigate the size and scope of the risk spreading of transgenic plants. In assessing the interrelation of GM plants and existing varieties, more detailed knowledge of the development of GM plants will provide a clearer, more reliable and precise directed activity in plant sciences.

Key words: *Genetic engineering, crop production, GM plants, biotechnology*

INTRODUCTION

Genetic manipulations have been commonly used throughout modern plant selection, including new combinations of genes, artificial manipulation of the number of chromosomes, the development of addition, substitution and breeds containing specific chromosomes, chemical and radioactive treatments aiming to induce mutations and chromosomal rearrangements, cell, tissue and embryo cultures, in vitro fertilization and fusion of protoplasts for detection of interspecies and genus hybridization. Integration of these technologies have made greatest contribution to genetic improvement in yield, adaptation to the environment (Dale, 2002), resistance to pests and parasites (Beringer, 2000), as well as improving the quality, that was permanently required by both food producers and consumers (Dunwell, 1998, Kuiper et al., 2001, Simic et al., 2005). Developments in molecular biology and genetic engineering enabled the efficient modification of cultivated plants (Daniell, 2002; Glisin, 2005). These technologies may

have an adverse impact on the environment (Conner et al., 2003), human health and increase poverty. During the forthcoming period, it is necessary to devote more attention to commercially and economically justified use of GM crops in the agricultural food production (Pretty, 2001, Nishiura et al., 2002; Boskovic et al., 2004a, 2010a, 2011, 2016). In this respect, linking science and politics should facilitate an overall assessment of the acceptability and spread of GM crops. There are still significant disagreements about the extent to which sustainability, globalization, ethics, and socio-economic approach should be parts of the risk assessment of GM crop use in most countries (Marshall, 2002; Marshall and Moonen, 2002, Perry et al., 2003; Simic et al., 2005).

PLANT PROTECTION AND GENETICALLY MODIFIED CROPS

Major plant protection considerations about GM crops (GMCs) are the possibilities of incorporation of resistance genes into insects, fungal and bacterial pathogens, viruses, emphasizing the long-term resistance as the main problem of (Bošković et al., 2000). Pest plant protection by genetic modification is demonstrated by the use of Bt toxins from *Bacillus thuringiensis*, successfully used as a spray for many years, that has been introduced in a number of plant species such as tomato, tobacco, cotton, etc. It has been proved that pea lecithin protects against insect attack in transgenic potato and tobacco. Trypsin inhibiting protein of fodder pea has also been used. GM plants provide a great possibility of benefits to the environment by reduced pesticide use by the development of pest resistance. However, these advantages can be quickly denied: there are possible strategies of exposure reduction of pests to transgenic products, thus lowering resistance level and applying restricted transgene action (Sharma et al., 2000; Bouchard et al., 2003). It has been shown that the application of transgenic resistance to viruses through indirect protection of *cap*-proteins is possible, and it can be used as a method for a wide range of viruses and hosts, e.g. expression of TMV *cap*-proteins for tobacco mosaic virus of potato and tomato resistant to PVX and PVY (Daniell, 2002). The use of genetic modification in fungal and bacterial plant pathogen control has also been developed. As with pests, the main problem is transient resistance and complex interactive pathogen-hosts relations. The improvement of such resistance was attempted to achieve by the use of GM plants. In addition, there is a need to develop multiple resistance strategies (i.e. *pyramiding resistance genes*) towards different virulence levels of plant parasites (Daniell, 2002, Bouchard et al., 2003; FAO, 2003; Bošković et al., 2004a; 2004b; Bošković and Bošković, 2009).

HORIZONTAL GENE TRANSFER (HTG)

Horizontal gene transfer (HTG) is a transfer of the genetic material between cells or genomes that belong to different species, during processes that differ from the common reproduction. During basic reproduction processes, genes are transferred vertically from parents to their offspring. Bacteria are known to participate in gene exchange between different species in nature. It is performed via three mechanisms: during conjugation when genetic material passes between cells, by transduction where the genetic material is transferred from one cell of infective viruses into another and by the transformation in which genetic material is taken directly from the cell or its environment (Daniell, 2002). For the successful horizontal gene transfer, the foreign genetic material must become the part of the cell genome, or be maintained stable

within the recipient cell. In certain cases, the foreign genetic material that enters the cell will be removed before it incorporates into the genome, especially if it originates from the different species. Under specific yet not fully clarified ecological conditions, foreign genetic material avoids the removal and incorporates into the genome. Horizontal gene flow is well known in bacteria, but also in some higher plants and animals. In its base, this process comprises the whole biosphere, and bacteria and viruses are intermediary in gene flow, gene pool, replication, and recombination (Conner et al., 2003; Snow et al., 2004). There are many potential pathways for horizontal gene transfer in plants and animals. As there are many viruses that infect plants and animals, transduction is expected to be the most frequent way of transfer. The latest investigations in gene therapy indicate the potential high importance of transformation for mammal cells, including humans. Direct transformation is not as significant for plant cells that have protective cell walls. However, soil bacteria that belong to the genus *Agrobacterium* are able to transfer T (tumor) segment of its induced tumor (Ti) plasmid into plant cells in the conjugation process. This Ti-DNK is widely used as a resource of gene transfer in plant genetic engineering. Foreign genetic material from insects and arthropods with strong mutagenic reactions can also be built-in into plant and animal cells. Additionally, bacterial pathogens that penetrate into plant and animal cells can take over foreign genetic material and carry it over in cells as vectors, probably to any kind of organism on the planet. The most significant barriers for horizontal gene flow are employed after penetration of the foreign genetic material into the cell. However, viruses and the other genetic parasites, such as plasmids and transposons, have special genetic signals and probably a structure that enables them to avoid annihilation. The genetic material of viruses is protected by the protein coat. They discard their protein coat when penetrating the cell, thus enabling the creation of numerous copies or pass directly into the cell genome. Plasmids are *free fractions* usually round in shape so that genetic material can finally be maintained in the cell separately from the cell genome. Transposons (*jumping genes*) are blocks of genetic material that have the ability to incorporate into the genome but can survive out of chromosomes, with or without the replication, and are kept within plasmids for further reproduction. Genes of genetic parasites such as viruses, plasmids, and transposons have a significantly higher ability to be successfully transferred into the cell genome. In such a manner they function as vectors for horizontal gene transfer regulated by inner characteristics of the organism, linked to the specific ecological conditions (Dale, 2002; Snow et al., 2004).

EVIDENCE OF HORIZONTAL TRANSGENIC DNA TRANSFER

It is believed that once incorporated transgenic DNA becomes stable. However, there is a body of evidence against this assumption. There are molecular data that prove structural stability of transgenic DNA, with regard to its location, point of penetration into genome and gene arrangement within the following generations. Transgenes can be either stabilized in successive generations or completely lost. Gene for herbicides tolerance incorporated into *Arabidopsis* with a vector can show up to 30 times stronger vector avoidance behavior, and it spreads as the identical gene obtained by mutagenesis. The results obtained suggest that horizontal gene transfer can occur via insects visiting plants for pollen and nectar, and that pollen can transfer transgenic DNA into the bee larvae gut (Thompson et al., 2003; Snow et al., 2004). It has been experimentally confirmed that the secondary horizontal transgene and gene resistance markers to antibiotics transfer in genetically modified plants, soil bacteria, and fungi. Transfer to

fungi was achieved by simple co-cultivation, while the transfer to bacteria by reisolation of transgene DNA or a whole DNK of a transgenic plant. Successful transfer of kanamycin resistance gene markers to the soil bacteria *Acinetobacter* was achieved by the use of extracted total DNA from homogenized plant leaf of the following transgenic plants: *Solanum tuberosum* (potato), *Nicotiana tabacum* (tobacco), *Beta vulgaris* (sugar beet), *Brasica napus* (canola) and *Lyopersicum esculentum* (tomato). It has been evaluated that about 2500 copies of kanamycin resistance genes are enough for the successful transformation of a bacteria, regardless if there exist 6 millions of folded strands) of the present plant DNA. However, natural conditions in the environment are widely unpredictable and some studies on synergetic effects could not be neglected in this case. Free transgenic DNA would be free in the rhizosphere around plant roots, which is a significant critical point in the environment. Some other scientists have found the horizontal transfer of the kanamycin resistance gene from transgenic DNK to *Acinobacter*, and positive results have been obtained by the use of homogenized 100 μ of the plant leaf. The biotechnology industry insists that the existence of horizontal gene flow in laboratory conditions does not mean that it occurs in natural environments. However, there is a body of evidence showing that it can and does occur in nature. Above all, genetic material taken from dead and living cells resists outer conditions, does not dissolve or devastating as it has been previously assumed. This indicates that sand, humus acid parts, and plant debris enable infections with more microorganisms in the soil. Bacterial transformation in soil via absorbed DNA into sandy clay has been confirmed by various experiments on microorganisms. In 1993 researchers in Germany have started a series of experiments with the rhyzomania resistant transgenic sugar beat plants (Pidgeon et al., 2001), that contained gene marker for kanamycin resistance, studying the stability of transgenic DNA and horizontal gene flow of transgenic DNA into the soil bacteria. Thus, horizontal gene flow represents a leading phenomenon that has occupied a significant place in the evolution of species, and it still does. All this suggests that natural horizontal gene flow presents a regulated process, limited by specific obstacles and mechanisms of rejection and inactivation of foreign genetic material. Genetic engineering created great variability of artificial construction for overcoming of barriers between all kinds of organisms and forestalling all genomes (Robinson et al., 2000; Glišin, 2005; Topisirović, 2005).

STRESS TOLERANCE AND PLANT RESISTANCE TO HERBICIDES (HRC)

Great efforts have been directed also toward the increase of crop tolerance toward herbicides, and it has been tested worldwide (Kwon et al., 2001). Herbicide tolerance can be achieved through the increase of protective mechanism, by reduction of herbicide uptake, by degradation or reduction of susceptibility (Caseley et al., 1991; Bradley et al., 2000; Konstantinović and Bošković, 2001). Herbicide tolerance genes are widely used as markers in transgene plant selection. If genes are of different tolerance to herbicides, developed or incorporated into the identical plant species, they could stipulate creation of weeds with multiple resistance genes (Colbach et al., 2001a; 2001b; Dewar et al., 2000; 2003; Desplanque et al., 2002; Richter, 2002a; 2002b). Numerous genes for tolerance to various environmental stress were inserted into plants. Greenhouse experiments showed that metallothioneins inserted in tobacco increased cadmium tolerance. Manipulation of genes that affect compatible solutions, such as betaine-aldehyde dehydrogenase, improve tolerance to salinity. The spread of crops tolerant to

drought and salinity could increase the range of wild relatives, making them more competitive, with consequential negative effects on ecosystems and agroecosystems (Beringer, 2000; Fagan et al., 2001; Marshall, 2001). Hybridization of plants resistant to herbicides (HRC) with populations of wild relatives makes these plants complicated for control, especially if they possess resistance to widely used herbicides (Altieri, 2000, FAO, 2003; Firban et al., 2003). Transgene plant resistance to herbicides (Kwon and Kim, 2001; Lutman and Berry, 2003) makes chemical control easier, above all because it includes compounds that are active to the very wide spectrum of weed species (Kim, 2001; Kudsk and Streibig, 2003).

QUALITY CHARACTERISTICS OF GM PLANTS

Transgene diversity is directed towards the improvement of products derived from cultivated plants. The effects of these changes on the whole plants are reflected in their metabolism and are of immediate interest to their commercial application. Some of these plants, especially commercially produced antisense" tomatoes" in the United States have been tested in field trials. The simplest change is probably the removal of the natural gene, for example, the „antisense“ tomato polygalacturonase gene. Other modifications include metabolic pathways changes via the introduction of various metabolic enzymes from other species, turning the natural path to a different end product. This can result in manufacturing plant products in different plant species, animal or bacterial products in plants or completely new products (Robinson et al., 2000; FAO, 2003). There are no *a priori* concerns about these types of GM plants because the modifications are very different. Metabolism modification in classical selection sometimes produces unexpected and undesired secondary effects. Unacceptably high levels of tannin were found in brown sorghum seed cultivated for resistance to birds. Therefore, random change of characteristics, such as the nutrient value of the main product or an increase in toxic secondary metabolites, should be carefully considered. It is necessary to collect as many biochemical and other data on transgenic plants, before they are widely accepted, in order to avoid unintended consequences that may become a problem (Gueritain et al., 2003). Adaptive effects of plants or hybrids with new characteristics must be studied case by case (Tester, 2001; Conner et al., 2003).

NON-CULTIVATED POPULATIONS OF TRANSGENIC PLANTS

The possibility of transgenic plant introduction will be highly dependent upon their adaptive capacity to the new environment. For many crops it is known that they form temporary non-cultivated populations, and these are species such as canola, lucerne, radish, carrot, rye, clover, sugar beet, chicory, beet, cabbage, of which some are native, whereas the others have probably been introduced for cultivation (Marlander et al., 2003). In some cultivated crops, i.e. rye, the difference between non-cultivated and the natural population is unclear, whereas in the case of some other species the colonization is not extensive and there probably has been no harmful influence on non-cultivated plant species. Possibilities of gene flow from the trial field with GM plants through pollen will depend upon sexual compatibility between GM crops and their wild relatives, and possibilities for pollination and obtaining of seed (Pretty, 2001). The occurrence frequency of such gene transfer will be influenced by important spatial isolation between GM crops and suitable recipients that depend upon the method of pollination,

wind or insects, isolation in time, i.e. flowering season. The experiments have been carried out in order to determine rates of cross-pollination between GM and non-GM potato planted at different spatial distances one from another (Hudson et al., 2001). The results were well harmonized, both showing that transgene movement outside the GM trial field has been negligible at distances less than 10 m, and low rates of cross-pollination that is usually present in potato have also been in accordance. In contrast, in canola seed the compatible inbreeding is present, it can produce huge seed quantities, it is pollinated by wind and insects. Pollination at huge distance happen probably due to the insects; air-born pollen can be found 30-50 m away of canola plants, but it is reduced with increasing distance (Bartsch et al., 1999). Darmency et al. (1998) have estimated the importance of pollination performed by honeybees in pollen dispersal from transgenic oilseed rape and concluded that although bees can travel up to 1-2 km from their hive, the food is carried from the vicinity of the hive. Experimental trials on fields that use GM or non-GM plants can provide useful data in regard to necessary isolation distances used to avoid the release of the transgene. However, trials on natural populations suggest that the situation could be more complex, and sub-classifications of local populations can strongly influence the transgene incorporation into wild populations (Dale, 2002). Interpretation of the results has also been complex, and Breeze et al., (1999) emphasize the prevailing importance of the calculation of dose changes in relation to distance from GM experimental field, over the absolute percentage of GM seeds at a given distance from the plot. However, gene flow can depend not only on crops but also on variety, location and season (Marshall et al., 2003). Experiments on gene flow in populations of wild radish suggest that the size of donor and recipient population plays a significant role in gene flow. In that way would huge pollen sources, such as the great spectrum of GM gene introduction have a significant influence on a small wild population of compatible plants. They also note that there are noticeable variations in evaluations of gene flow, probably due to local-positional and pollinator effect. Other researches have found similar effects in *Cucurbita* (Moonen and Marshall, 2001) and rice (Song et al., 2002). Further work on gene spreading in populations can be necessary during the assessment of the potential transgene dispersal. Such assessments become particularly important in analyzing the possibilities of transgene transfer from cultivated crops to related species (Darmency et al., 1998; Rongli et al., 2000).

GENE TRANSFER FROM CROPS TO WEEDS

Global group for wheat has identified three crops that have sexually compatible weed relatives for which it is probable that it will be subject to the gene transfer in agricultural systems. Breeding via pollination and production of fertile hybrids varies from case to case. Even when chosen characteristics have a positive advantage, introgression of new characteristics into the existing weed population is still possible (Konstantinović and Bošković, 2001). The risk of environmental damage depends upon weed habitat. In weed-crops complexes (Hauser and Ostergård, 1998) analyzed in these studies, where the habitats of weed relatives are restricted to agricultural systems, it is not possible that new trait endangers natural ecosystems (Creswell et al., 2002).

ROLE OF AGRICULTURAL WEED ECOLOGY IN ACCIDENTAL HYBRIDIZATION

Weed ecology and evolutionary biology are of high importance in the assessment of perspective of the accidental flow of resistant transgene to harmful organisms in the population of agricultural weeds (Bradley et al., 2000; Marshall et al., 2003). The model for accidental transgene flow has three phases that lead to creating a widely distributed transgene carrying weed populations. The first phase presents a hybridization between weed and transgenic crops. The second phase is the occurrence of introgression and adaptation processes with evolutionary mechanisms that improve inadaptable traits in hybrid products of an earlier generation. As a consequence, weeds transfer resistance transgene to harmful organisms (Buckelew et al., 2000) and due to that have a high adaptation level for certain agroecosystems. Finally, the process of dispersion and dissemination of these new weeds in nature, together with local adaptation of various conditions is important during the transition of a sufficiently wide area. Weed ecology and evolutionary biology are important for understanding the interaction between the three previously described phases. The real scientific data on weed ecology are still lacking. During the few last decades, scientists focus on herbicide weed control. Amongst these prevailing studies, ecological research and especially theoretical description have been neglected (Altieri, 2000; Konstantinović and Bošković, 2001; Conner et al., 2003). Hybridization between transgenic or conventional plant species and sexually compatible relatives occurs in many crops and produces new forms of weeds in obtained populations. In numerous papers, this hybridization has detailed description and it can be expected that transgenes will transfer even over great spatial and significant obstacles of genetic incompatibilities (Perry, 2002). In some systems, the accidental transfer of transgenes by hybridizations seems unavoidable. However, in other cases, it is not clear if hybridization is a proportionally limiting phase in the transfer of transgenes. It is the assumption that hybridization can even be proportionally limiting in some circumstances, i.e. when hybridization occurs over significant obstacles of incompatibility. Aspects of weed ecology that can influence hybridization levels in these situations include weed cropping systems and effects of spatial and timely distribution of weeds in several phases. The selection system between weeds in field agroecosystems of crops is a mixed system of fertilization in which occurs inbreeding and cross-fertilization (outbreeding), although the other reproductive system has also been known. Therefore, the widely distributed systems of weed selection enable hybridization, but such fertilizations must happen during a significant level of inbreeding. For selection systems and other aspects of genetic systems and reproductive ecology, that influence the hybridization levels it is known to vary within and between weed populations. For instance, *Datura stramonium* populations in Northern Carolina have flowers that open to pollinators and exhibit approximately 10% outbreeding level. Quite the contrary, certain populations are exclusively self-pollinating, with flowers that do not open to the pollinators. In some cases this variation refers to the adaptation of the selection system after great expansion (Cresswell et al., 2002), the however attitude of the pollinators can vary in many ways, even geographically. These aspects of reproduction, therefore, should not be considered as permanent characteristics in the frame of weed species (Pidgeon et al., 2001; Moyes et al., 2002). Spatial weed distribution can strongly influence hybridization weed-plant species. First, many weeds have uneven distribution in fields, and in some papers, it is described that uneven distribution in some species has a certain level of temporary stability (Perry et al., 2003). Uneven distribution can be caused by edaphic factors or by persistent effects of

high seed production. In the frame of a field, uneven weed distribution can reduce hybridization weed-crop. The occurrence of weeds in plant populations of adequate density, with a small proportion of single weeds at the ends of these plant populations, is limited populations proportionally to hybridization. More homogeneous and uneven distribution can significantly favor to considerably higher levels of cross-fertilization. In locally isolated individuals can be present higher quantities of crop pollen due to changes in the movement of pollinators as a function of local density. Weed density can have a contra-effect to the hybridization levels when the plant serves as the female parent. In this case, high densities can favor hybridization with the advantage of the achievement of great local weed pollen densities, and homogenous weed density can reduce plant hybridization. Weed distribution in a wide area around field crop agrosystems also has potential importance in determining the level of weed-crop hybridization. If conditions allow stabilization of weeds in areas that are not under crops in the region, then many small, isolated populations can exist (Pierre et al., 2003). Weed-crop hybridization can occur at higher levels in these populations than in populations in the field due to a number of reasons. For example, because of the differences in reproduction in many weed species, the flowering can occur over a long period of time during the growing season of a given species. Seed from the commercial crop fields, pollinated by wind or insects, scatter for more than 1 km outside the field. Therefore the crop pollen can be expected to reach the weed populations outside the agricultural fields within this distance. When considering the population of weeds in a particular environment, the proportion of intense flowering and the presence of pollen of cultivated plants can significantly extend the range of opportunities for hybridization in many weed-crop systems. Many weeds are highly variable. During certain years weather factors can lead to failure in controlling weeds in the wider region, resulting in a high density of weeds in some areas. Weed density varies on a regional basis because of the interaction between the biology of weeds and regionally dependent different ways of weeds distribution and other factors of the breeding system. Both forms of variation can result in a significant increase in the absolute number of hybridization cases. The level of hybridization is affected by the density of weeds, dependent variations in the behavior of pollinators or spatial distribution. For example, increased the local presence of the species may enable them to settle border habitats in agricultural areas, which are densely populated, probably increasing the likelihood of hybridization. Therefore, levels of hybridization can vary significantly over the years and a number of weed species (Rongnli et al., 2000; Richter et al., 2002a, 2002b).

TRANSGENE INTROGRESSION AND WEED ADAPTATION

The evolutionary process that follows hybridization will certainly be influenced by many ecological weed properties in agrosystems of cultivated plants (Kwon et al., 2001). The nature of these systems seems prevalent only for strongly expressed factors of weed population regulation (Pidgeon et al., 2001), compared with the majority of annual plant populations, which are short-lived and inhabit other types of ecosystems. This can facilitate transgene introgression even if the hybrids and starting backcrossed generation have a low level of adaptive features in comparison to weeds that do not carry transgenes (Fagan et al., 2003; Kim, 2001). Weed populations are frequently small and sometimes temporary, so the effects of selection, migration and random genetic changes will influence the evolution of introgression. Seed ecology is of primary

importance for weed survival, therefore, the effects of transgenes on other genes in plant species and seed ecology will probably induce strong selective effects in these genes. Introgression of genes that improve the adaptation of weeds to these predominant selective factors can significantly increase the average adaptability of the weed population. Exchanges between adaptations to different limiting factors that result from introgression of a single gene can also be minimal. The best example is the evolution of herbicide resistance in weeds (Kwon and Kim, 2001; Lutman and Berry, 2003; Pierre et al., 2003). The occurrence of herbicide resistance often significantly increases the average survival and population growth of weed. Herbicide-resistant mutation can have high absolute adaptability, despite basic functional damages that are caused by pleiotropic effects of resistance mutations. This shows how selection can favor a mutant that exceeds limiting factors. The other line of evidence rises from multiple examples of increased distribution and density of weeds from field trials over hybridization (Perry et al., 2003). Finally, many cases of the basic increase in distribution and density of certain weeds follow moderate changes in cropping systems, providing the additional proof that many weed communities are regulated with several strong factors (Rognli et al., 2000). If correct, this assumption suggests that the adaptation of weeds can rely upon randomly transferred transgenes after hybridization and is made easier by the biological uniformity of the actual field crop ecosystems. Weeds can demand relatively narrow evolutionary movement, as it is a break of linkages toward undesirable characteristics of plant species in accordance with adaptation to wide-area (Saeglitz et al., 2000). One criterion for assessment of transgene spread into weed population is that the survival rate of the hybrid weed-crop that carries transgene should be higher than the adaptability of non-hybrid weeds. This criterion can be much easier met in temporary ecosystems of field crops than in the majority of others. Therefore, the transfer of transgenes can be a fast process. Even hybrids with very low adaptability and early back-crossings can survive in the agrosystem's inadequate densities. There is a possibility for introgression and adaptation that make their survival easier. These assumptions can be applied to accidental transgene transfer that influences tolerance to abiotic factors. However, weed populations in wider areas can be restricted by one biotic factor adaptation to which would provide basic survival increase (Streinbrecher, 1996; Snow et al., 2004). The other characteristics of weed ecology that probably influence the adaptation of the crop-weed hybrid are the frequency of occurrence of low efficient population size and high levels of self-fertilization, especially during the colonization process (Colbach et al., 2001a; 2001b). Small population sizes cause random changes in genetic composition. These mechanisms can act at a genetic base produced by hybridization, producing a number of genetically differentiated small populations from genetically different back-crossed weed populations. This implies that adaptation in weed populations that contain random transgenes probably will be influenced by selection and random genetic change. The presence of both factors enables the occurrence of evolutionary processes that do not occur when selection is the dominant evolutionary mechanism. Specifically, adaptive effects of transgene combinations, other crop genes, and weed genes can be more adaptable in expression with mutual action of random genetic exchanges in regard to the selection that acts alone (Conner et al., 2003). These mechanisms can be especially expressed when weed populations have high levels of weakening and repeated colonization, forming ecological and genetic metapopulations. Although, it is still not clear if weeds in agriculture have metapopulation structure, the occurrence of such

structures, in combination with small-sized population and altering selection pressures create convenient conditions for equilibrium processes. However, the effect of these processes can itself be unpredictable due to geographic variations in population structure in some weed species due to the cropping system, local adaptation after colonization, the time span from colonization and hybridization with related taxonomic categories (Dale, 2002; Marshall and Moonen, 2002). The final dimension of weed ecology in relation to adaptation after hybridization is the ecology of seeds. The ability to maintain seed stability in the soil, together with effective spreading and the ability of fast and efficient reproduction are the most important features in the spread of weeds in crop field agro-ecosystems (May, 2003). The weed populations dynamics have shown that seed demography (i.e. survival and germination levels) greatly affects growth levels of weed populations. The longevity of weed seeds varies considerably between species. Many agricultural activities affect the demographics of seed amounts in the soil, preventing germination or otherwise increasing the seed extinction levels. These factors include stubble crops, tillage, and residue burning, that can affect the seed directly or indirectly, via effects on the animals that feed on seeds and pathogens. Germination and dormancy of weed seeds are important for survival in any given developmental system (Marshall, 2002). If the transgenes affect the weed ecology, the caused effects are likely to act as factors dominant for survival (Sharma et al., 2000) hence leading to weed infestation. Similarly, the unadjusted effects on the ecology of seed may be the primary mechanism by which non-transgenic crops reduce the adaptation after hybridization. There are several other aspects of the ecology of weed seed, which affect the adaptation phase. The first is the well-known effect of dormancy, where the genotypes of plant weeds that grow out of environmental conditions, can again be separated for the storage during adverse conditions periods. Therefore, the seed populations are affected by the genetic variability of weed populations. Also, the seed population, as a form of temporary expansion of genotypes allows weeds to be tested over a wider range of conditions that would not be possible otherwise. This effect can significantly increase the possibility that a weed carrying accidentally transposed transgene appears in an environment to which it was adapted. Molecular and biochemical data on homology levels between species of crops and their wild relative's worldwide indicate gene introgression from crops into populations of wild relatives (Nakayama and Yamaguchi, 2002), in various plant species including maize, melon, carrot, sugar beet and rice. The possibility of transgene introgression into wild populations, during time within some species and in some geographic areas, would be very high. Hybrids between crops and their wild relatives most probably occur in crops with low adaptation grown in their region of origin. Minimal divergence and maximum exposure from the other relatives should be provided. Level of transgene introgression from crops into wild species is influenced by factors such as overlapping of the flowering period with wild species, and capability of hybrid to back-cross with wild relatives within the population (Darmency et al., 1998). This suggests that gene transfer possibility is especially significant in tropic areas because many important crop species evolutionary originate from these areas. Isolation distance or incompatible flowering time can be realized through the use of lines - barriers with different species, guard-lines with the same species and application of genetic engineering causing male sterility (Saeglitz et al., 2000) that might contribute to genetic isolation. Evidence suggests that isolation itself could reduce the degree of introgression due to pollinating insects (Sharma et al., 2000; Bouchard et al., 2003), while the use of barriers or guard plants could be useful for limiting the spread of GM

pollen from the experimental field. Other proposed methods of genetic isolation include the selection for increased feeding in GM crops or reduced sexual compatibility with wild relatives. The use of suicide genes was also considered, however, such measures involve significant limitations in the application of GM plants.

SPATIAL SEED DISPERSION AND WEED COLONIZATION

Efficient spatial dispersion of seed is considered a primary feature of weeds (Kudsk and Streibig, 2003) and it is expected that weed ecology influences the fate of the accidentally transferred spread transgenes in a number of ways. On the field scale, simulation modeling indicates that high levels of weed seed dispersal generally greatly increase the population of weeds (Pidgeon et al., 2001). For most weeds in field crops dispersion is determined by the interaction of characteristics of weeds and human activities, such as contaminated seeds of cultivated plants, equipment, water for irrigation and seed transmission. When human activities become the main vectors of the spread of weed seed, those dispersions are difficult to characterize, because of the geographical variations of the processing system. As a result, the maximum spreading distances are not known in most cases. On a wider scale, many cases of rapid expansion of the sub-continental weed species are known. Weed species were found to become abundant over large areas west USA (Rognli et al., 2000), due to the changes in farming methods that improve their abundance, such as *Aegilops cylindrica*, sexually compatible weed in wheat. Weeds resistant to herbicides spread over hundreds of kilometers of road shoulders in less than a decade. These observations suggest that weed populations of road shoulders of other non-field habitats may be important for the spatial spread of weeds, recognizing the importance of weed ecology in agricultural areas and the accidental spread of the transgenes (Altieri, 2000; Marshall, 2002). Herbicide-resistant crops can be agronomically harmful if the resistant species germinates before the sown plants germinate, or if the herbicide-resistant plant occurs as a weed on another field (Tyystjarvi, 2009). Weeds can have significantly higher adaptability over large spatial areas of the transferred transgene. In theory, the resulting spatial homogeneity of suitable habitat (Perry, 2002; Conner et al., 2003) and the absence of the need for local adaptation accelerates the rapid expansion of the colonizing organisms. Therefore, the ecology of weed spread and population regulation in agro-ecosystems and agricultural areas seems to enable large and rapid expansion of the adapted weeds. The weed properties that affect their spreading (e.g. seed size, shape, similarity to crop seeds, etc.) should be considered adaptive characteristics that are probably the result of a strong selection. The effects of crop genes and ecological expansion can adversely affect the adaptation of hybrids. The weed spread may have an evolutionary role. In a small basic population, it can cause an adaptive process, which does not occur in large populations. In the increasing balancing process, the spread of weeds has an important role in evolution, moving into small populations in other areas and affecting changes in other populations. The weed ecology in cultivation systems can facilitate the random transgene transition, allowing the survival of weed-crop hybrids that are not adjusted, compared to wild-type weeds, in a series of adaptability components. This probably happens when the hybrids and next-generation back cross, carrying the appropriate transgene adaptive value. Ecology of seed, the expansion of some basic characteristics responsible for the weed adaptability and the level of population growth is not widely accepted parameters. Effects arising from transgenic and other crop genes will greatly

influence the adaptation of weed-crop hybrids and back cross generations. Most of the main weed species show intensive spatial and temporal variations in reproduction, ecology and seed spread on several scales. This variation has both genetic and environmental causes (FAO, 2003; Snow et al., 2004). The populations of agricultural weeds are widespread in agricultural regions, including many populations that appear out of farmland. Specific ecology of these populations may affect all phases of the random transgene transfer (Dewar et al., 2003).

THE INFLUENCE OF GM PLANTS ON BIODIVERSITY

One of the limitations for the introduction of GM plants into the environment is the concern about the adverse effects of those plants on biodiversity, including the possibility of its devastation. Fear of loss of biodiversity (Anon, 1994) is an important basis for opposition to genetic modification and GM plants by several influential environmental protection research groups from around the world. The influence of GM plants on biodiversity is a complex and complicated problem. Scientific studies and discussions have been directed towards understanding whether GM crops have an impact on biodiversity and defining their qualitative and quantitative differences from commercial crops. Biodiversity and agriculture are strongly interrelated since biodiversity is crucial for agriculture, while agriculture can contribute to the sustainable use of biodiversity (De Jaramillo, 2009). Biodiversity is very important for the survival and maintenance of global planetary conditions, providing the aesthetic, scientific, cultural and other values. The general value of the world's biodiversity is estimated to the amount of about 33 trillion \$ per year (Costanza et al., 1997; Bošković et al., 2012a). Regarding the multidimensional complexity of biodiversity concept and taking into account the significance of the technological development of GM plants, further studies that will clarify this interdependence are needed. In a broader sense, they will rely upon the social, economic and political context of genetic modification application that will determine the risks or potential advantages of GM plants to biodiversity.

TRANSGENES AND MONITORING

The genetic modification itself does not substantially change anything, but the adaptability of GM plants will depend on the transgene effects and the impact of a particular stage in the transgene development. For example, the length of the vitality of seeds and seedlings stabilization can be an especially important modification of oil crop seed. Both may be more important for the stabilization of annual plants, in relation to the characteristics that affect the survival of adult plants or fertility. The effects of transgenes within natural populations of ecologically important plants, which may have a greater effect on the species that carry the transgene were also examined (Song et al., 2002). If transgenes were giving a selective advantage to a wild type, it might become dominant, which would result in a reduction in the natural variation (Steinbrecher, 1996). Certain characteristics of transgenic genetic modification may present an advantage in some environments, e.g., tolerance to salinity, drought, cold and pest resistance. A wide variety of existing genes for transgenes transfer (Watkinson et al., 2000) and the changes feasible by molecular techniques, make this technology fundamentally different from traditional methods of selection. Ecological monitoring of GM crops in complex ecosystems is needed even after commercialization. This

complexity varies from year to year and indicates indirect biotic effects. As laboratory and field, experiments cannot sufficiently repeat all interactions that occur in one ecosystem, the only way for evaluation of the full level of ecological effects of GM plants is monitoring in natural ecosystems. Some of these effects cannot be predicted in advance, so ecological monitoring will be needed to reveal and differentiate existing ecological influences (Altieri, 2000; Dale, 2002; Bošković and Isajev, 2007). Environmental monitoring is very expensive, so the information obtained should be used within the clear system of adaptive management. Such management includes repeated cycles of firmly set rules within specially designed programs, the use, evaluation and estimation of the monitoring as a whole (Tester, 2001; Snow et al., 2004). A more significant problem in GM plant monitoring is the lack of adaptive management systems specifically developed for this application. To control the evolution of resistance to pathogens and pests, the frequency of resistance should be monitored in the field, with the additional research in order to set standards for monitoring and determining appropriate management that can clarify the problem (Boskovic et al., 2010b). Monitoring of new GM plants will need to be evaluated by the diverse group of scientists and scientific disciplines including agriculture, forestry, the ecology of wetlands, entomology, pathology, etc. In the future scientists and technological advances will continue to expand opportunities for artificial design and construction of plant organisms. Genomics and bioinformatics will facilitate the identification of commercially important genes that can potentially be transmitted between species. Environmentalists will more significantly contribute to the wider debate on the public contribution to the prevention of risks of those innovations.

CONCLUSIONS

Biotechnology alongside the introduction of genetically modified (GM) crops is constantly providing new opportunities for increasing crop productivity and tackling problems in agriculture, such as diseases, pests and weeds, abiotic stress and nutritional limitations of stopple food crops. Crops possessing new traits enabling the use of pharmaceutical products are also being generated. As GM crops are being introduced into various locations with different ecosystems, agriculture, biodiversity and agriculture practice, a scientifically based understanding of the environmental effects of GM crops cultivation would assist decision makers worldwide in ensuring environmental safety and sustainability. The main important environmental assessment of GM crops deals with their putative invasiveness, vertical and/or horizontal gene flow, effects on biodiversity and the impact on other products. These investigations are all highly interdisciplinary and complex. This paper deals with some of the most important problems related to entering GM crops into the environment, such as plant protection, ecological effects of HRCs, gene flow, biodiversity, stress, ecological risks of Bt crops, effects on soil ecosystems, etc. There is a clear need to further assess the severity, magnitude, and scope of risks associated with the massive field deployment of transgenic crops. When assessing GMC inter-relation with the existing cultivars, an increased knowledge base underpinning the development of GMC will provide greater confidence in plant science while assessing the risks and benefits of releasing such crops.

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